1	EXPLORING THE MICROBIAL COMMUNITY INHABITING THE PHOSPHOGYPSUM
2	STACKS OF HUELVA (SW, SPAIN) BY A HIGH THROUGHPUT 16S/18S rDNA
3	SEQUENCING APPROACH.
4	
5	Authors:
6	Patricia Gómez-Villegas ¹ , José Luis Guerrero ² , Miguel Pérez-Rodriguez ³ , Juan Pedro
7	Bolivar ² , Antonio Morillo ¹ , Javier Vigara ¹ , Rosa Léon ^{1,*}
8	
9	Affiliation:
10	¹ Laboratory of Biochemistry, Center for Natural Resources, Health and Environment
11	(RENSMA). University of Huelva, Avda. de las Fuerzas Armadas s/n, 21071 Huelva,
12	Spain.
13	² Department of Integrated Sciences, Center for Natural Resources, Health and
14	Environment (RENSMA), University of Huelva, Avda. de las Fuerzas Armadas s/n,
15	21071 Huelva, Spain
16	³ Department of Cell Biology, Physiology and Immunology, University of Córdoba,
17	Campus de Excelencia Internacional Agroalimentario CeiA3, Córdoba, España.
18	
19	*Corresponding author:
20	Rosa León
21	Laboratory of Biochemistry, Center for Natural Resources, Health and Environment
22	(RENSMA). University of Huelva, Avda. Fuerzas Armadas s/n.
23	21071-Huelva, Spain.
24	e-mail: rleon@uhu.es
25	Tel: +34959219951

26 ABSTRACT

27 Around 100 Mt of phosphogypsum (PG) have been deposited in large stacks on the salt 28 marshes of the Tinto River estuary in Huelva (SW Spain), covering about 1000 ha. These 29 stacks contain extremely acidic water (pH < 2) with high concentrations of pollutants 30 which can cause emissions into their surroundings, generating important environmental concerns. Despite many chemical, geological or hydrological studies have been 31 conducted to characterize the PG stacks of Huelva, the microbial community inhabiting 32 33 this extreme environment remains unexplored. Using a 16S/18S-rRNA- high throughput sequencing approach, we have uncovered the main taxonomic groups able to live in the 34 acidic metal-contaminated water, which is in direct contact with the PG. In addition, the 35 physicochemical characteristics of the water sampled have been analyzed. These 36 37 studies have revealed that the most abundant bacteria found in two different leachate samples of the PG stacks belong to the genera Acidiphilium, Pseudomonas, 38 Leptoprilium, Acidithrix or Acidithibacillus, which in total represent around 50% of the 39 total bacterial community. These iron-oxidizing and/or reducing bacteria are typical 40 41 acidophilic genera usually found in acid mine drainage (AMD) environments. Biodiversity of eukaryotes in PG water is lower than that of prokaryotes, especially in the water 42 collected from the perimeter channel that surrounds the PG stacks, where the pH 43 reaches a value of 1.5 and the activity concentrations exceed 300 Bq L⁻¹ for ²³⁸U or 20 44 Bg L⁻¹ for ²¹⁰Po, values which are from four to five orders of magnitude higher than those 45 usually found in unperturbed surface waters. Even so, an unexpected diversity of algae, 46 47 fungi and ciliates have been found in the PG stacks of Huelva, where chlorophyte microalgae and basidiomycetes fungi are the most abundant eukaryotes. Additional 48 49 bioinformatics tools have been used to perform a functional analysis and predict the most 50 common metabolic pathways in the PG microbiota.

51

52

53 Highlights

54	The extreme conditions of these PG stacks hide an unexpected microbial diversity								
55	The prokaryotic profile includes acidophilic bacteria found in AMD environments								
56	Archaea, cyanobacteria and sulfate-reducing bacteria are practically absent								
57	High presence of microalgae in the piezometer denotes seawater influence								
58	• The eukaryotic population in the PG is different from that of the AMD environments								
59									
60									
61									
62	Keywords: 16S/18S rRNA; Acid mine drainage; High throughput sequencing;								
63	Metabarcoding; Pollutants; Phosphogypsum; Radionuclides.								
64									

65 **1. INTRODUCTION**

The phosphoric acid is mainly obtained from the phosphate rock (PR) by treatment with dilute sulfuric acid in the so-called "wet process". In this process, a by-product called phosphogypsum (PG), which basically consists of gypsum (CaSO₄·2H₂O) but also contains significant concentration of harmful pollutants, is generated. About 300 Mt yr⁻¹ of PG are worldwide produced, being most of it dumped in big waste stockpiles usually placed in coastal zones.

72 Near the city of Huelva, in the southwest of Spain, around 100 Mt of PG have been 73 generated from 1968 to 2010 by the fertilizer manufacturing companies operating in the 74 area for more than forty years. This huge amount of PG was transported as an aqueous 75 slurry and stockpiled, without any prior treatment or insulation layer, over the salt marsh 76 sediments located in the common estuary formed by the confluence of the Tinto and Odiel Rivers, covering more than 1000 ha and generating a significant environmental 77 impact (Guerrero et al. 2021a). Until 1998, about 20% of the PG generated was directly 78 79 discharged into the Odiel River channel, while the remaining product was mixed with 80 seawater and pumped into the stacks, where PG was decanted, and the polluted acidic 81 seawater was released into the Tinto River channel without any treatment. After 1998, all the PG generated until the end of the phosphoric acid production activity, in December 82 83 2010, was pumped with freshwater in a closed circuit and placed into the piles to comply with the regulations of the OSPAR Convention (OSPAR, 2002). 84

The PG stacks of Huelva are currently divided in four zones (Fig. 1). Zones 1 and 4, with a total extension of around $5.5 \cdot 10^6 \text{ m}^2$ and located at the north and south of the affected area, respectively, are already considered as restored. However, polluted edge outflows from these two zones have been detected reaching the water of the Tinto estuary (Pérez-López et al., 2016). In Zones 2 and 3, the uncovered PG is directly exposed to weathering and contains superficial acidic water ponds, shown in green color in Fig. 1. Depth and extension of these ponds are progressively decreasing by evaporation during the

ongoing restoration of the area. In zone 2, about 25 Mt of PG are accumulated covering 92 a surface of around 2.5.10⁶ m² and forming a pyramidal pile of up to 30 m high, 93 94 surrounded by a network of perimeter channels to collect the leachates from the stack. The PG stacks of Huelva have been exhaustively studied from the geochemical point of 95 view with particular attention to the pollution pathways. Aspects such as the mobility of 96 97 heavy metals and radionuclides contaminants present in the PG (Pérez-López et al., 98 2018; Guerrero et al., 2020) or the composition of the leachates released from the stacks 99 (Pérez-López et al., 2016; Papaslioti et al., 2018; Guerrero et al., 2021a) have received special attention. However, to date, the microbial community that inhabits the PG piles 100 101 of Huelva has not been explored at all. The extremely toxic composition of the PG, 102 besides the adverse conditions for life (high solar irradiance and salinity, low water 103 content, and extremely low pH) have probably discouraged research in this sense.

104 PG contains impurities such as fluoride, diluted phosphoric acid, heavy metals, or 105 radionuclides coming from the original phosphate rocks as natural contaminants (Bolívar 106 et al., 2009; Guerrero et al., 2020), which increase the potential environmental impact of 107 PG and limit the proliferation of microorganisms. U-series radionuclides, heavy metals, 108 such as Cd or Ni, and metalloids as As are naturally present in most phosphate ores, and they are transferred into the PG during the production of the phosphoric acid (Bolivar 109 110 et al., 2009). The high mobility of these pollutants at the low pH values of the PG 111 leachates causes their easy dissolution. Therefore, the outflows coming from the PG 112 piles or their pore waters contain very high concentrations of heavy metals and 113 radionuclides, which can reach concentrations from 3 to 5 orders of magnitude higher than those observed in unperturbed waters, as we have recently demonstrated (Guerrero 114 115 et al., 2021a)..The presence of such pollutants, the acidic nature of PG and the high proportion of compounds with biocide activity make the physiological conditions of PG 116 highly inhospitable. However, there are many examples of extreme environments where 117 microorganisms are able to proliferate, becoming adapted to acidic or basic pH values, 118

high concentrations of NaCl, heavy metals, organic solvents, or other xenobiotics, and
demonstrating the adaptation ability of microbial life to colonize a wide range of extreme
ecological niches, and the PG stacks of Huelva are not an exception (Merino et al., 2019).

122 There are PG deposits at several locations worldwide, especially near the main phosphoric acid production plants, in China, USA, Tunisia, or Morocco. However, the 123 microbial species able to thrive under these adverse conditions and flourish in contact 124 125 with the PG have remained practically unexplored. The few studies published in relation to the microbiology of PG have been carried out in the deposits of Gafsa and Sfax, in 126 127 Tunisia (Mefteh et al., 2019; Trifi et al., 2020). These studies were focused on the 128 microbiota associated with the PG sediments and with certain plant species able to grow 129 in the vicinity of the PG. However, the microbial community present in the leachates of 130 these PG stacks or in the PG deposits of other locations has not been investigated to 131 date.

133 2. MATERIALS AND METHODS

134 **2.1. Sampling**

The two PG leachate samples used in this study were collected at the beginning of July, 135 136 before the total loss of water by evaporation that usually takes place at the end of the annual hydrological cycle. The first water sample was taken from the perimeter drainage 137 channel (Fig. 1, code "PC") that surrounds the Zone 2 of the stacks. The second one 138 139 was taken from a piezometer (Fig. 1, code "PZ") that collects the groundwater from the 140 south border of the same zone, around 3-4 m in depth. The UTM coordinates (WGS84 datum) for the sampling points in the channel and the piezometer are 29 S 684521, 141 4123390, and 29 S 684536, 4123295, respectively. Seawater, sampled near the city of 142 Huelva, and water of the Tinto River, collected at the lower course of the river (near the 143 144 town of Niebla) have also been included for comparison. The water samples were collected by an electrical pump with a flow rate of 4000 L h⁻¹ in sterile 25 L plastic tanks 145 146 and transported to the lab for their immediate processing.

147

148 **2.2. Determination of physicochemical parameters of the leachates**

Physicochemical parameters as electrical conductivity (EC), pH, oxidation-reduction potential (ORP), and temperature (T), were determined *in situ* by a Hanna HI98195 portable multimeter, with a HI7698195 multiparameter pH/EC/T probe. The pH/ORP sensor used (ref. HI7698194-1) contains a glass pH sensing tip combined with an Ag/AgCI reference electrode in gel electrolyte. The instruments were calibrated before sampling, and the ORP was corrected to obtain the potential relative to the hydrogen electrode (Eh).

156

157 **2.3. Chemical and radiochemical analysis of the leachates**

158 To determine the metallic composition of the water, the samples were digested with HCL at 150°C during 30 min and analyzed by Inductively Coupled Plasma Optical Emission 159 160 Spectroscopy (ICP-OES) using an Agilent 5110 spectrometer, equipped with a SeaSpray nebulizer. The argon line was used as an internal standard f. The quality 161 control (QC) was implemented by the measurement of Certified Reference Materials 162 (CRMs). To determine the anionic composition, the water samples were filtered through 163 164 a 0.45 µm filter and analyzed by ionic chromatography in a Methrom 882 Compact IC plus instrument, equipped with an 858Pro auto-sampler, a cationic suppressor, and a 165 Metrosep A-Supp 5-150/4-0 column. The mobile phase consisted of sodium carbonate 166 (0.339 g L⁻¹) and sodium bicarbonate (0.252 g L⁻¹). Flow rate: 1 mL min⁻¹. Injected 167 168 volume: 20 µL⁻¹. All measurements were done in triplicate.

169 Natural radionuclide concentrations in the collected leachates were determined by a sequential extraction technique based on the use of tributyl phosphate (TBP), 170 subsequent electrodeposition onto stainless-steel disc (U-Th-Ra isotopes), and self-171 deposition onto silver discs for the case of ²¹⁰Po. The radioactive sources were counted 172 by α-particle spectrometry using ion-implanted silicon detectors, with a 25% absolute 173 174 efficiency. The QC for alpha-particle measurements was conducted by participating in annual international proficiency tests (International Atomic Energy Agency [IAEA] and 175 176 the Spanish Nuclear Safety Council [CSN]), and by measuring both a blank and CRMs 177 (IAEA-434) every set of six samples (Bolívar et al., 2009).

178

179 2.4. DNA isolation

For genomic DNA isolation, 10 L of PG water samples, were filtered through a 0.7 μm glass fiber filter (Whatman, GF/G), and the retained biomass was later eluted by backflushing with some additional sample water. Finally, the concentrated samples were centrifuged at 12,000 xg and the biomass obtained was used for genomic DNA extraction

with the GeneJET Genomic Purification kit (Thermo Fisher Scientific, Waltham, MA,
USA), following the manufacturer's instructions.

186

187 **2.5. Library preparation and DNA sequencing**

Prokaryotic and eukaryotic rDNA amplicons were obtained by PCR with the Phusion® 188 High-Fidelity PCR Master Mix (New England Biolabs, MA, USA), using the genomic DNA 189 previously isolated as a template and the primer sets: 341F/806R, to amplify the 190 191 hypervariable V3-V4 16S rDNA region; and 1380F/1510R, to amplify the V9 18S rDNA 192 region (supplementary material, Table S1). The PCR products were cleaned up using 193 the Qiagen Gel Extraction Kit (Qiagen, Germany) and pooled to generate two libraries, 194 one prokaryotic and one eukaryotic, using NEBNext® UltraTM DNA Library Prep Kit for 195 Illumina and quantified accurately with the Qubit and Q-PCR. Library pools were 196 sequenced in the Illumine MiSeq platform using the Illumina MiSeq Reagent kit V2x 250bp, to generate paired-end raw reads. 197

198

199 2.6. Bioinformatic data processing and OTUs clustering

Bioinformatic analysis of the data was carried out using QIIME2 v2020.8 (Bolyen et al., 201 2019). Briefly raw data were demultiplexed, using the q2-demux plugin, and filtered by trimming and truncating low-quality regions, eliminating replicated reads, and filtering chimeras with DADA2 (via q2-dada2) (Callahan.et., 2016). The reads were then clustered in operational taxonomic units (OTUs) using a 97% similarity cutoff, with the *de novo* clustering method (via q2-vsearch) from VSEARCH (Rognes et al., 2016).

206

207 **2.7. Taxonomic classification and Functional analysis**

208 OTUs were classified at each taxonomic rank using the q2-feature-classifier plugin (via 209 classify-sklearn method) and the SILVA database (Quast et al., 2013) with two different 210 pre-trained classifiers, specially curated for 16SV3V4 and 18SV9 regions. Annotation was performed with a 0.7 threshold. The taxonomic information retrieved for the best hits 211 212 was used for the annotation at different taxonomic levels (kingdom, phylum, class, order, family, genus, and species, if possible). For the analysis of the species diversity, the 213 214 alpha-rarefaction tool was used via q2-diversity to generate a rarefaction curve as 215 previously described (Gómez-Villegas et al., 2018).

Finally, the inferential functional analysis was performed using PICRUSt2 (Douglas et al., 2020). The entire PICRUSt2 pipeline was used to generate metagenome predictions
from 16S rRNA amplicon through sequence placement, and pathway-level prediction by
predicting EC numbers, KEGG orthology terms (K numbers), and MetaCyc pathways.
KEGG pathway maps were obtained from the generated set of K numbers using the
KEGG Reconstruct tool (update: July 1, 2021). Venn Diagrams were generated using
the VennDiagram package (v.1.6.20) for RStudio (v1.3.1093).

3. RESULTS AND DISCUSSION

225

226 **3.1. Hydrochemical characterization of the leachates**

227 The values of the measured physicochemical parameters, the concentration of the most 228 significant elements and anions, as well as the activity concentration of the main natural 229 radionuclides in the two water samples collected from the PG stacks of Huelva, are 230 displayed in Table 1. The values corresponding to the two reference samples (seawater 231 and Tinto River water) have also been included in this table. While the physicochemical 232 parameters and the concentration of dissolved species in seawater remain approximately stable over time, the Tinto River shows a strong variability, for this reason, 233 234 the mean values for a hydrological year were included in Table 1.

235 The pH measured in the perimeter channel water was extremely acidic, with apH value 236 around 1.6. The Eh indicates the oxidizing conditions of the channel with a value of 674 237 mV. The EC showed a value of 42 mS cm⁻¹ which indicates a high concentration of 238 dissolved ions. The content of contaminants such as As, Cd, or Ni was also very high, 239 with values of 33, 9.8, and 6.3 mg L⁻¹, respectively, which is between 45 fold (for Ni) and 240 150 fold (for As) the values observed in the water of the Tinto River, which has been 241 included as a reference. The concentrations of other metals, such as Cu, Co, or Fe were of the same order or even lower than in the Tinto River, which is highly influenced by the 242 243 acid mine drainages. Other hazard contaminants such as Hg or Se were undetectable. The extremely high activity concentration of ²³⁸U-series radionuclides at this sampling 244 point is noteworthy. The values exceed 300 Bq L⁻¹ for ²³⁸U or 20 Bq L⁻¹ for ²¹⁰Po, being 245 from four to five orders of magnitude higher than those observed in unperturbed 246 freshwater and seawater. 247

The stacks of Huelva stopped receiving PG in December 2010, however maintenance and restoring works are ongoing and the area counts with several piezometers were built

250 along zones 2 and 3 of the stacks to monitor the phreatic level and to allow the collection 251 of groundwater samples. The location of the sampled piezometer is shown in Fig. 1. It has a total depth of 4.8 m and collects leachates through its last perforated meter. In the 252 253 point where the piezometer is located, there is 35 cm of restored soil over the PG, which 254 comprises a layer of about 5 m thick. This was our second sampling point, which allowed us to obtain underground polluted water from the bottom of the PG material. The pH of 255 256 the piezometer water was 1.7, an extremely low value, but slightly higher than that 257 observed in the water of the perimeter channel. The Eh denotes significant oxidizing 258 conditions, but the measured value (480 mV) was significantly lower than the obtained 259 in the channel, due to the fact that this sample has a groundwater source. Besides, the 260 EC value in the piezometer is 53.6 mS cm⁻¹, higher than the value at the channel., The 261 high EC value in the piezometer, very similar to the value of the seawater (Table 1), is 262 due to the influence of the seawater used for transporting the PG until 1997. The 263 concentrations for As, Cd, or Ni in the PZ, with values of 9.2, 2.5, and 1.8 mg L⁻¹, 264 respectively, are much higher than in the Tinto River, but not as high as in the perimeter 265 channel. The concentration of As in the piezometer is almost 50-fold the concentration in the Tinto River, however, this is around three times lower than the value measured in 266 267 the perimeter channel. Similarly, the concentration of Cd in the PZ is almost 30 times the 268 concentration in the Tinto River and four times lower than in the PC. In general, it can be 269 confirmed that the concentration of all the tested elements was lower in the piezometer than in the perimeter channel, excepting the cations K, Mg, and Na or the anion Cl, which 270 showed higher concentration values in the piezometer due to a clear influence of the 271 seawater. The activity concentration of the analyzed natural radionuclides was also very 272 high at this point, but one order of magnitude lower than in the perimeter channel for the 273 U- and Th-isotopes. The concentration of fluoride, a well-known biocide anion, is 274 particularly high in the perimeter channel water, with a value of 1.4 10³ mg L⁻¹. The 275 276 concentration of fluoride in the piezometer water was 85 mg L⁻¹ and undetectable in the Tinto River. 277

All these data indicate that the acidity, the content of toxic heavy metals, natural radionuclides, and fluoride are higher in the perimeter channel than in the undergrown water collected in the piezometer, which, on the other hand, presents higher electrical conductivity and higher salinity due to a clear influence of the seawater. However, in both cases, the low pH values and the high content of toxic metals will condition the microbiota which will have to be specially adapted to thrive with these harsh conditions

284

3.2. High throughput sequencing of the rRNA markers encoding genes and estimation of biodiversity in the PG water samples

287 The reads obtained from the high throughput sequencing of the 16S rRNA libraries 288 generated for both PG water samples were denoised, filtered to eliminate chimeras and 289 PCR artifacts, and clustered into 680 and 596 taxonomic units for the perimeter channel 290 and the piezometer water samples, respectively. Sequencing of the 18S rRNA libraries, on the other hand, yield 38 and 186 OTUs, for the perimeter channel and the piezometer 291 292 water samples, after processing by the same bioinformatics pipeline. The number of sequences and the mean quality score for each sample are shown in Table 2, while 293 294 rarefaction curves are in Supplementary material (Fig. S1).

It is noteworthy that the volume of water processed to obtain enough genomic DNA was about 5 fold higher in the PG water samples than in the Tinto River or seawater samples. Moreover, the number of raw reads and the number of non-chimeric merged inputs obtained is around one order of magnitude lower in the PG water samples than in the Tinto River or seawater samples. This indicates a lower biomass content in the PG water samples in comparison with the biomass found in the seawater or the Tinto River water, as was expected for the harsh living conditions that prevail in that environment.

The Shannon-Weiner biodiversity index (H) for the whole prokaryotic population, in both the perimeter channel and in the piezometer, are shown in Table 2 and compared with

the biodiversity index of the Tinto River or the Sea water. The biodiversity index for the
prokaryotic community is higher for the PG water samples, with values of 7.7 and 7.5,
than for the control Sea and Tinto River water samples.

Similar analysis done for the 18S rRNA libraries revealed that the Shannon biodiversity 307 index in the eukaryotic community is 1.5, in the case of perimeter channel water, and 308 309 4.5, in the water collected in the piezometer. Although the biodiversity index for the 310 eukaryotic population of the piezometer water is similar to that observed in seawater 311 (4.5), the value obtained for the perimeter channel is lower (1.5), indicating a more 312 reduced biodiversity within the eukaryotic community of the perimeter channel water. In 313 the Tinto River, the Shannon index for eukaryotes (6.3) is higher than that for the 314 prokaryotic population (5.1). In agreement with different previous studies, which have 315 repeatedly confirmed the paradoxical high eukaryotic diversity in the Tinto River, that 316 was even superior to that of prokaryotes (Zettler et al. 2002; Gadanho, et al., 2006; Aguilera 2013). 317

The extreme physicochemical parameters of the PG water suppose an inconvenience for the development of high concentrations of microorganisms, however, do not impede the existence of a large variety of species, especially of the prokaryotic kingdom. The extremely low pH value and the high redox potential of the water from the perimeter channel seem to affect more drastically to the eukaryotic community, which exhibited a very low biodiversity index (1.5) and presented a low number of OTUs.

324

325 3.3. Prokaryotic population in the water samples collected from the PG stacks

Analyzing the data obtained from sequencing the 16S library, we can conclude that in the PG water, from either the piezometer or the perimeter channel water, the most represented sequences correspond to the phylum Proteobacteria, which supposes 62% of the total OTUs in the perimeter channel and 67% of the piezometer, followed by the

330 phyla Actinobacteriota and Nitrospirota, which represent around 15% and 5%, 331 respectively, in both PG water samples. These phyla comprise bacteria of very different 332 characteristics, however, it is possible to observe that the phyla profile in the PG water 333 differs significantly from that of the seawater, where Proteobacteria are much more 334 abundant (90%). This profile is also very different from that of the Tinto River, where the 335 phylum Nitrospirota represents 23% of the total (Fig. 2).

Focussing at lower taxonomic levels, the most abundant OTUs in both PG water samples correspond to the classes Gammaproteobacteria and Alphaproteobacteria, which are also the most represented in the seawater and the Tinto River. However, the classes Leptospirillia and Acidimicrobiia which are very abundant in the perimeter channel with percentages of 23% and 7%, respectively, are poorly represented in the underground water of the piezometer. The classes Actinobacteriae and Bacteroidia, by contrast, are more abundant in the piezometer (around 3%) than in the perimeter channel.

343 If we go down in the taxonomic scale and look at the genus level, Acidiphilium is the 344 most abundant one, representing 10.4% of total OTUs in the perimeter channel and 345 about 11.3% in the piezometer. Followed by the genera Pseudomonas, Leptospririllum, 346 Acidithrix, Acidithiobacillus, Stenotrophomonas, and Ferrovum, which suppose between 347 4 and 5%. of the total OTUs. All these genera count with extremophilic representatives 348 and many of them have been found to be present in AMD environments, where the most 349 abundant taxa include the genera Acidiphilium, Acidisphaera, Acidithiobacillus, and Leptospirillum (López-Archilla 2001; Lukhele et al. 2020). Most of these genera coincide 350 with the main genera represented in our sample of the Tinto River, in which 25% and 351 352 20% of the OTUs correspond to the genera Acidiphilium and Leptospirillum. The genera Acidithrix (6.1%), Acidithiobacillus (10.7%), and Ferrovum (11.1%) are also present at 353 high percentages in our sample from the Tinto River. 354

355 *Acidiphilium* is an acidophilic genus, usually found in AMD environments This versatile 356 genus has been proposed to have arisen in ancestral non-extreme conditions and have

357 got adapted to harsh acidic environments by the acquisition of new functional features through the evolution. The trait for this adaptive evolution can be observed in the different 358 359 genomes available for species of the genus, which contain a repertory of genes, acquired 360 by horizontal gene transfer, to cope with metal and osmotic stress or to metabolize different nutrients (Li et al., 2020). In our study, the most represented species of the 361 genus are Acidiphilium cryptum (100% sequence homology) and Acidiphilium rubrum 362 363 (99.5% sequence homology), species also found in AMD environments in Canada (Auld 364 et al., 2013) and Turkey (Aytar et al., 2014). From a metabolic point of view, most Acidiphilium sp. are aerobic acidophilic heterotrophs that utilize organic compounds to 365 366 obtain carbon and energy, although many can obtain some extra energy from the 367 mixotrophic oxidation of ferrous iron. In addition, some members of the genus have been described to be able to reduce Fe³⁺, respiring iron under anaerobic or microaerophilic 368 conditions or to perform photosynthetic CO₂ assimilation (Kisková et al. 2018). 369

370 On the other hand, the genera Acidithiobacillus and Leptospririllum comprise a series of 371 obligate or facultative chemolithoautotrophs widespread found in ADM sites. The genus Acidithiobacillus is an important representative of the sulfur-oxidizing bacteria. The 372 373 species of this genus are strict autotrophic Gram-negative bacteria with a versatile 374 metabolism that can fix both carbon and nitrogen from the atmosphere, using ferrous 375 iron or reduced sulfur compounds as a primary source of energy and generating the 376 corresponding oxidized ferric or sulfur species, contributing in this way to the sulfur 377 biogeochemical cycle. They can grow aerobically or anaerobically, using ferric iron as 378 the final electron acceptor. The most abundant species of this genus in our samples were 379 Acidithiobacillus ferriphilus and A. ferrooxidans, which are important contributors to the 380 sulfur biogeochemical cycle. A low representation of Acidithiobacillus thiooxidans, 381 formerly known as *Thiobacillus thiooxidans*, has been found in the PG samples of Huelva in this study. This species is one of the best-known iron-oxidizing acidophiles, very 382

abundant in the acidic environment of the Rinto River and essential for the recovery of
 copper by bioleaching (González-Toril et al., 2003).

385 The genus Leptospirillum includes different iron-oxidizing species, which role in the 386 acidic environment has gained an increasing acknowledgment (Méndez-García et al., 387 2015). In our study, Leptospirillum ferrooxidans, which represents 3.4% and 3.7% of the total OTUs in the perimeter channel and the piezometer, respectively; and Leptospirillum 388 ferrodiazotrophum, which suppose around 1% of the total OTUs in both locations, are 389 390 the most represented species of the genus. These species are present a much higher percentage in the Tinto River water (16.3% for L. ferroxidans) and are practically non-391 392 existent in the seawater.

Other acidophilic genera found in the Huelva PG samples are Acidithrix, and Ferrovum. 393 394 The first species of the genus Acidithrix described was Acidithrix ferrooxidans, isolated 395 in north Wales from an acid mine drainage site and classified as a new heterotrophic acidophilic species. Acidithrix, as other acidophiles, can use Fe²⁺ (but not sulfur) as an 396 397 inorganic electron donor and an organic carbon source, getting extra energy from this 398 mixotrophic nutrition. Under the limitation of oxygen species of this genus have been described to reduce ferric iron. The reads assigned to Aciditrhix in our study only share 399 400 89% of sequence identity with A. ferroxidans, however, they have a 100% of sequence 401 identity with other uncultured Aciditrix genera identified in metagenomic surveys in 402 locations as distant as the acidic hot springs of the Colombian Andes (Bohorquez et al., 403 2012), the extreme acidic Pitlakes from the Iberian Pyrite Belt (Santofimia et al., 2013) 404 or polymetallic deposits in Russia (Gavrilov et al., 2019). The species of the genus 405 Ferrovum are iron oxidizers chemolithoautotrophs that pair Fe(II) oxidation with the 406 fixation of carbon dioxide as a carbon source although there is a high metabolic diversity 407 within the genus. We have found in the PG a high percentage of OTUs (4 and 4,6%) with high sequence homology with uncultured Ferrovum species found in acidic 408 409 environments.

410 The genus *Ferrovum* belongs to the order Burkholderia. Some strains of this order, such as Burkholderia fungorum- or Paraburkholderia fungorum strain Rifles, have been 411 412 described to respire uranium, anaerobically reducing U (VI) into U (IV) for their growth (Agarwal et al., 2019). Several OTUs assigned to uncultured species of the 413 414 Burkholderiaceae family have been found in our studies. The sequences corresponding to these OTUs showed an identity of only 90% with the corresponding sequence of P. 415 416 fungorum, however, the high concentration of uranium in the perimeter channel water 417 (Table 1) makes probable the existence in this sample of new linages related or not with 418 P. fungrum, with the ability to reduce uranium.

419 Other abundant genera are Pseudomonas and Stenotrophomonas, which represented 420 about 4.9 and 4.4%, respectively, of the total 16SrRNA OTUs obtained from the 421 perimeter channel; and about 2% and 5.1% in the case of the piezometer water. The 422 genus Stenotrophomonas is widely spread in aqueous environments. Their ability to form 423 biofilms and their versatility can explain the capacity to colonize extreme environments 424 of some species of the genus. Recently, the tolerance of this Fe–Mn oxidizing bacterial 425 genus to high concentrations of iron and its ability to treat AMD effluents has been 426 demonstrated (Hou et al., 2020). However, this is not a typical acidophilic genus, and it 427 is also present at a high concentration in the sea water control sample. *Pseudomonas* is other ubiquitous versatile genus with species adapted to many different extreme 428 429 conditions, including acidic environments (Kisková et al., 2018).

It is noteworthy that the prokaryotic population present in the PG water comprises almost exclusively bacteria (99.8 %), being the presence of archaea practically null. The acidophilic or sulfur-chemolithoautotrophic archaea described to date are related to hot acid springs and are mainly restricted to the order *Sulfolobales* (Ghosh and Dam, 2009), which is not represented in our study.

Zouch and coworkers found the development of sulfate-reducing bacteria (SRB), such
as *Desulfobacter or Desulfovibrio*, in enrichment cultures inoculated with sediments

collected from the vicinity of Stockpiled Sfax Phosphogypsum in Tunisia (Zouch et al.,
2017), but we found not a trace of these bacteria in the Odiel PG water. Neither,
cyanobacteria, which are among the most widespread primary producers both in the
oceans and in the freshwater, are present in the PG leachates, although several genera
of this group appear in the seawater sample.

442 Comparison of the microbiota found in the Huelva PG leachates samples with that of other similar PG sites is impossible because there are no similar studies. The few studies 443 444 carried out to identify the microorganisms in the PG areas of Tunisia were not focused 445 on the water but on the sediments or the scarce vegetation that proliferate in the zone. 446 Mefteh and coworkers (Mefteh et al., 2019) concluded analyzing the sediments of the 447 Tunisian PG of Sfax that the dominant phylum was Proteobacteria with more than 40% 448 of the total prokaryotic sequences, being the most abundant genera Staphylococcus, Bacillus, Pseudomonas, and Acidithiobacillus, although important differences were 449 found depending on the bioinformatics pipelines and the databases used for the 450 taxonomic assignment. 451

Trifi and coworkers (Trifi et al., 2020), on the other hand, explored the same PG Tunisian location, focusing on the microbiota present in the PG sediments and associated with four higher plants growing in the contaminated PG vicinity. They found that the most abundant phyla in the sediments were Firmicutes (39.3%), Proteobacteria (24.2%), and Actinobacteria (20.4%) and the dominant genera were *Bacillus* (20.7%) and *Enterococcus* (17.6%).

None of these results coincide with the microbial profile that we have found in the PG stacks of Huelva. However, as we have already pointed, the Tunisian study was done in the sediments while we explored the microbial community in the water reservoirs of the PG. In addition, the PG water in Huelva can be highly influenced by the acid mine drainage discharged into the Tinto River, making the microbial community of the Huelva PG stacks a unique microbiota.

464

465 **3.4. Eukaryotic population in the water samples collected from the PG stacks**

466 The most abundant eukaryotic phylum in the perimeter channel is Basidiomycota, which 467 constitutes 47.3% of the total OTUs obtained (Fig. 3). Specifically, the fungal groups 468 Filobasidiales and Agaricales, were the most abundant orders within this phylum. A more detailed analysis reveals that the fungal species found share a high sequence identity 469 470 (98%) with the yeast Solicoccozyma sp. and the widespread fungus Schizophyllum 471 radiatum, which are versatile basidiomycetes with genetic and biotechnological interest 472 (Li et al., 2020). However, no acidophilic strains of these species have been described until now. The percentage of fungi (including yeast) within the eukaryotic community of 473 474 the piezometer leachates is low in comparison with the perimeter channel. The phyla 475 Basidiomycota and Ascomycota only represent 4% and 1.1%, respectively, of the 476 eukaryotic OTUs in the piezometer.

Although several yeast species have been isolated from extreme aquatic environments, 477 such a Cryptococcus, Rhodotorula, or Lecytophora, found in the Sao Domingos mine in 478 479 Portugal (Gadanho et al., 2006) and in the Tinto River in Spain (Aguilera 2013), 480 acidophilic strains of the yeast Soliccoyma have not been reported in these locations. 481 Another acidic environment in which the eukaryotic heterotrophic community has been deeply studied is the Rio Agrio, a naturally acidic river in the vicinity of a volcanic area in 482 483 the Argentine Patagonia. In this acidic location, the main acidophilic yeasts described 484 belong, as in the areas influenced by acidic mine drainages, to the species Cryptococcus 485 and *Rhodotorula* (Russo et al., 2008). The absence of these typically acidophilic fungal 486 species in the PG leachates is noteworthy. Similarly, different genera of extremophilic filamentous fungi have been found in hyperacidic environments (López-Archilla et al 487 488 2004). However, to our knowledge, the existence of acidophilic strains of the fungi Schizophyllum in other locations, apart from the PG of Huelva, have not been yet 489 490 reported. The pH values in the PG piezometer and perimeter channel are around 1 and

491 1.2 units lower than the medium pH value of the Tinto River water. This suggests that
492 the pH of the PG leachates must be under the threshold pH tolerable by the typically
493 acidophilic fungal species found in other acidic water bodies.

494 Regarding photosynthetic eukaryotes, the phylum Chlorophyta is the most represented, 495 supposing 18% of the eukaryotes in the perimeter channel and 62.1% in the piezometer 496 (Fig. 3). Chlorophyta is also the most abundant phylum in the seawater and in the Tinto 497 River, where it reaches 40.7% and 81.8% of total OTUs, respectively. The phyla 498 Ocrhophyta (6.6%), Dinoflagellata (1.7%), Diatomea (1.4%), Protalveolata (1.3%), and 499 Cryptophyceae (1%) are also present at significant percentages in the piezometer 500 leachates. Instead, in the perimeter channel the only microalgal phylum of importance, 501 apart from the Chlorophyta, is the Dinoflagellata, which represents 11.1% of the total 502 eukaryotics OTUs. Curiously, the diversity of photosynthetic eukaryotes contrasts the absence of cyanobacteria in the PG leachates and in acidic environments in general. 503

504 The phylum Chlorophyta is a highly diverse group of algae closely related to higher 505 plants. Most of the species of this phylum are freshwater, however, there are important marine genera. In the Piezometer water, families of this phylum, such as 506 507 Treboxinophyceae (48%) and Chlorophyceae (3.2%) are abundant, while in the 508 perimeter channel the most represented chlorophyte group is Chlorodendrophyceae, 509 which includes the genus Tetraselmis. In fact, OTUs which share 90-93% sequence 510 identity with several species included in the database as Tetraselmis sp. or unclassified Prasinophycea are highly represented in the perimeter channel. 511

512 Comparison with the eukaryotes present in other PG deposits is not possible, because 513 none of the studies done in other PG affected areas, such as those realized in Tunisian 514 PG, have investigated the eukaryotic population (Mefteh et al. 2019; Trifi et al. 2020). 515 Once again, the most related environments are those affected by acid mine drainages 516 or volcanic activity. The eukaryotic community of the Tinto River is one the best studied 517 in an extremely acidic, heavy metal-contaminated habitat, and microalgae occurrence

and diversity in this acidic river have been well studied. Genera of the phylum Chlorophyta, such as *Chlamydomonas, Chlorella*, or *Dunaliella* have been detected in the Tinto River (Zettler et al., 2002; Aguilera, 2013). Some examples of these genera have been reported to possess high metal tolerance (León-Vaz et al., 2021) and many of them have shown their ability to easily evolve into acidic and heavy metal adapted species (Zettler et al., 2002).

524 Diatoms of the phylum Ochrophyta, such as the genus *Pinnularia*, have been previously 525 reported to be present in the Rio Tinto waters (Aguilera, 2013), however, we have not 526 found this genus in our survey. The most abundant Ochrophyta representatives in the 527 piezometer are the genera Poteriospumella (1.1%), Nanochlorpsis (1.4%), and an uncertain genus of the family Chromulinales (3.5%). Whithin the Dinoflagellata phylum, 528 529 the genera Gyrodinium (7.7%) and Alexandrium (0.4%) are the most represented in the perimeter channel, and Alexandrium (1.1%) the most abundant in the piezometer. This 530 last genus is also significant in the seawater (8% of total eukaryotic OTUs). Both 531 dinoflagellates have widespread marine distribution. Some species of the genus 532 Alexandrium are known as usual components of toxic algal blooms, but no acidophilic or 533 534 acidotolerant species of this genus have been described so far. Other non-535 photosynthetic eukaryotes, which do not belong to the group of the fungi, have appeared also in our study of the Huelva PG. The phylum Ciliophora is found in the perimeter 536 537 channel (12.3%) and in the piezometer (2.2%) at a high percentage. The phylum 538 Cercozoa, on the other hand, is present at percentages of 1.2 and 0.2% in the piezometer 539 and the perimeter channel, respectively. The sequences retrieved in this study share a 540 high percentage of identity with other uncultured ciliates and ameboflagellates, however, 541 there is not enough information for their assignation to a specific genus. Previous studies 542 describe the existence of ciliates, amoebas and other protist predators in the Tinto River 543 (Zettler et al. 2002).

Additionally, traces of other eukaryotes, including, arthropods, nematodes, higher plants, or rotifers have been detected in the PG waters. Monogonontos, a sub-clase of rotifers that includes some acidophilic species, is an example present at a considerable percentage in the piezometer water (4%), but not in the perimeter channel. This rotifer class was discovered in highly acidic mining lakes (Ersabek et al., 2011).

549 To our knowledge, there has not been any previous description of the eukaryotic 550 community able to live at PG sites in any world location. However, there are several 551 studies that report the presence of eukaryotic extremophiles in acidic volcanic environments (Russo et al 2008) or acidic mine drainages sites (Gadanho et al., 2006). 552 553 One of the best-studied acidic environments is the Rio Tinto, where an unexpectedly high diversity of eukaryotes was found (Zettler et al., 2002; Aguilera, 2013). However, 554 555 differently from what was observed in the bacterial profile, the eukaryotic community found in the PG differs significantly from that inhabiting the acidic mine drainage waters. 556 It seems that not all the eukaryotes able to dominate the extremely acidic, heavy-metal-557 558 contaminated mine environments are able to tolerate the additional harsh conditions of 559 the PG water or the even lower pH values.

The influence of the seawater in the piezometer sample is clear, not only because of its physicochemical parameters, but also due to the high percentage of marine microalgae and a much lower representation of freshwater fungi and ciliates, which are the main eukaryotic group in the perimeter channel water.

564

565 **3.5. Functional modelling analysis**

This research reveals the extraordinary variety of prokaryotic and eukaryotic microorganisms inhabiting the leachates obtained from Huelva PG. This population includes autotrophs and heterotrophs which assimilate inorganic or organic carbon compounds, respectively. Includes aerobic species which use oxygen as the final

570 acceptor of electrons in the respiratory electronic chain, and a series of versatile anaerobic species able to replace oxygen by ferric iron or other oxidized species as 571 572 electron acceptors. Many acidophilic bacteria are facultative anaerobic that can adapt 573 their respiratory metabolism to diverse environmental conditions or use the reduction of ferric iron to obtain extra energy. Regarding the autotrophs, besides the oxygenic 574 575 photosynthetic microalgae, a series of acidic chemolitotrophes able to fix CO₂ and obtain 576 energy from the oxidation of ferrous iron, sulfide, or other reduced sulfur compounds, 577 have been found in this study. Fig. 4 schematically represents the main pathways of this 578 complex panorama, citing some of the main representatives of each metabolic group.

To obtain further insight into the functional characteristics of this microbial population the sequences retrieved from the metaxonomic study on the basis of the 16S rRNA marker gene were analyzed with the use of the PICRUSt2 software and the functional gene databases Kyoto Encyclopedia of Genes and Genomes (KEGG) or *MetaCyc*. This analysis has allowed us to predict the abundance of gene families and obtain ontology pathway predictions in the studied PG leachates and in the seawater or the Tinto River samples (Fig. S2)

586 For both, the perimeter channel and the piezometer samples, the main functions predicted using the PICRUSt algorithm can be classified within the categories: Metabolic 587 pathways, Genetic information processing, Environmental information processing, and 588 589 Cellular processes. A more detailed insight within the most represented category reveals that the carbohydrate metabolism, with 900 or 912 orthologous genes (OG) in the PZ 590 591 and the PC respectively, the amino acid metabolism, with 613 and 620 OG in the PZ and 592 the PC respectively, or the energy metabolism with 473 and 500 OG in the PZ and the 593 PC, respectively, are the most represented metabolic pathways (Fig. S2A). Among the 594 most represented metabolic routes in the PG leachates, as well as in seawater and the 595 Tinto River, those related to methane metabolism, oxidative phosphorylation or amino 596 sugar, and nucleotide sugar metabolism stand out (Fig. S2B).

597 Comparative functional analysis between orthologous genes in the two studied PG 598 leachates shows that there are 5937 common orthologous groups, while 292 and 346 599 are unique for the PZ and the PC, respectively. Regarding the functional categories, about 389 had representation in both PG leachates and only 11 of them were exclusively 600 601 found in the PZ or the PC. Many of these genes and functional pathways are also found in the Tinto River and the seawater, although there are two functional pathways from PC 602 603 the and one from the PZ which are unique to these locations, and have been found neither in the seawater nor in the Tinto River, such as the biosynthesis of 604 605 mannosylglycerate (PWY-5656) or the degradation of creatine (PWY-4722) are 606 exclusively found in the PZ or the PC, but not in the seawater or the Tinto River (Fig.5). 607 Mannosylglycerate (MG) is an osmoregulatory solute typical of some halophilic 608 organisms and creatine is a natural amino acid that can act as a source of nitrogen and 609 carbon under certain circumstances. The presence of these pathways can be related to the existence of extreme conditions and low content of nutrients. 610

611

612 **4. CONCLUSIONS**

613 The leachates released from the PG stacks of Huelva constitute a unique environment of extreme acidity and high concentration of heavy metals and radionuclides, which has 614 615 not been characterized from the microbiological point of view although these microorganisms can play an important role in the dynamics of the polluting compounds 616 of the PG. Our study demonstrates for the first time that, in spite of the extreme 617 physicochemical conditions, a high diversity of prokaryotic and eukaryotic 618 microorganisms can thrive in the two locations studied, the perimeter channel that 619 620 surrounds the PG stacks and the underground water collected in the piezometers settled 621 at the borders of the stacks.

The prokaryotic profile resulted to be quite similar in the two leachates studied. The most abundant bacteria correspond to acidophilic genera, usually found in environments influenced by AMD, such as the Tinto River. Chlorophyte microalgae and Basidiomycota fungi, are the most abundant eukaryotic groups. This metabarcoding survey has also made clear the practically inexistence of representatives of the domain Archaea, as well as the null existence of species from the prokaryotic group of photosynthetic cyanobacteria or from the group of the sulfate-reducing bacteria.

Acidic and oxidizing conditions are more extreme in the perimeter channel than in the piezometer, where conductivity and salinity are higher, confirming the higher influence of the seawater in this sample. The influence of seawater in the PZ is also evidenced in the composition of the eukaryotic community. While the eukaryotic population of the PC water is dominated by the fungi, which represents 47% of the eukaryotes, the PZ has a much more important contribution of marine microalgae, which reach 62% of the total eukaryotes.

This study shows that the high content of pollutants of the PG influences the microbial diversity in the leachates. The information about the composition of the microbial community of the PG can shed light on the possible role that these microorganisms play in the dynamics of the contaminating compounds of the PG and can contribute to the obtaining of new strains with biotechnological applications.

Acknowledgments: The authors thank to Fertiberia S.A. for its support in obtaining the
water samples used in this study. P. Gómez-Villegas acknowledges the financial support
of the University of Huelva (EPIT 2016-17).

644

645 Funding: This research was funded by University of Huelva and the Operative FEDER 646 Program-Andalucía 2014-2020 (UHU-1257518 and UHU-1255876); The SUBV. COOP.ALENTEJO-ALGARVE-ANDALUCIA 647 2021; The European Regional Development Fund through the Agencia Estatal de Investigación (research grant PID 648 2019-110438RB-C22 and PID2020-116461RB-C21) and the Andalucian government 649 650 (I+D+i-JA-PAIDI-Retos projects 2020- PY20_00728 and PY20_00096).

651

Figure captions

652

Fig.1. Map of the Huelva phosphogypsum stacks with definition of the different zones and location of the sampling points: piezometer (PZ) and perimeter channel (PC).

656

Fig. 2. Relative abundance of prokaryotic OTUs at various taxonomic levels.

658 Perimeter channel (PC), piezometer (PZ), sea water (SW) and Tinto River (TR)

659

660 **Fig. 3**. Relative abundance of eukaryotic OTUs at various taxonomic levels

661

Fig. 4. Main metabolic processes occurring in the PG leachates. Some significant
 representative species are included. Light grey metabolic pathways and /or
 species have not been found in these samples. See the text for a more complete
 description

666

Fig. 5. Venn diagrams of the number of shared and unique functionally annotated
orthologous genes (A) and metabolic pathways (B) in the perimeter channel (PC),
the piezometer (PZ), the sea water (SW) and the Tinto River (TR) obtained using
MetaCyc database for functional comparision. (C) MetaCyc codes for unique
pathways in each location

672

673

674 Supplementary material

675

Table S1. Primer sequences used to target the corresponding hypervariableregions for the construction of the libraries.

Fig. S1. Rarefaction curves. The Observed features and the Shannon index
values for the 16S and the 18S data are plotted against the sequencing depth,
which referees to the product of the number of sequences by their length in
nucleotides.

682

Fig. S2. KEGG Mapper Reconstruction Result. The number of functionally annotated genes (K terms) for each of category of pathway (A) and for the majoritary metabolic pathaways (B) obtained from Picrust analysis (<u>https://github.com/picrust/picrust2</u>) of the piezometer (PZ), the perimeter cannel

687 (PC), the seawater (SW) and Tinto River (TR) samples is shown

688

689

690

Table 1. Physicochemical parameters, concentration of the most significant694elements and ions (mg L-1), and activity concentration (Bq L-1) of the main695radionuclides in the water samples collected from the perimeter channel (PC) and696the piezometer (PZ) at the PG stacks of Huelva (SW, Spain). Values for the697seawater (close to Huelva City) and Tinto River (near Niebla Town) from previous698studies have been included for comparison.

Variable	PC	PZ	Seawater ^a	Tinto River ^b	
pН	1.56	1.7	1.7 7.78		
T (°C)	25.5	21	12.8	-	
Eh (mV)	674	480	465	689	
EC (mS cm ⁻¹)	41.8	53.6	61.5	2.4	
As	32.8	9.2	< 0.002	0.224	
Cd	9.81	2.5	<0.002	0.089	
Co	0.8	0.2	<0.002	0.54	
Cu	9.8	3.1	0.0035	18.9	
Fe	104	34.9	0.0081	144	
Hg	<0.01	<0.01	-	<0.01	
Mn	14.5	7.6	<0.002	8.6	
Ni	6.3	1.8	<0.002	0.14	
Pb	0.5	0.1	0.003	0.097	
S	1722	1246	1040	1550	
Sb	0.9	0.11	0.011	<0.1	
Se	<0.1	<0.1	-	<0.1	
Zn	74.3	20.8	0.050	20.7	
Ca	1250	832	441	82	
K	459	561	440	5.1	
Mg	667	730	1440	85	
Na	4027	9536	11700	46	
F ⁻	1377	85	111	2.9	
Cl⁻	6778	16619	23563	72	
SO4 ²⁻	17077	3231	3288	4810	
²³⁸ U	302	24	0.042	0.057	
²³⁴ U	302	23	0.045	0.112	
²³⁰ Th	12	0.62	<0.0001	0.044	
²³² Th	0.12	0.03	<0.0001	0.023	
²²⁶ Ra	0.73	0.62	-	-	
²¹⁰ Po	22	43	0.0036	0.011	

^aGuerrero et al., 2021a. ^bGuerrero et al., 2021b

	Raw	After	Mean quality			Observed	Shann
Sample	Reads	denoising	inputs	Q ₃₀ (%)	Q Score	OTUs	on
							Index
PC_16S	19129	14995	10555	94.60	≥ 36	680	7.7
PC_18S	10067	8864	6076	93.36	≥ 36	38	1.5
PZ_16S	17129	13236	9409	94.54	≥ 36	596	7.5
PZ_18S	30362	21890	15080	89.74	≥ 36	186	4.5
TR_16S	189164	175352	145395	95.12	≥ 36	348	5.1
TR_18S	160136	126280	103365	90.99	≥ 36	133	6.3
SW_16S	206161	190181	171677	95.17	≥ 36	838	6.4
SW_18S	203677	172627	111906	92.24	≥ 36	399	4.5

703 **Table 2.** Sequence data statistics

704

The number of raw sequences, sequences after denoising, non-chimeric merged inputs, and Observed Operational Taxonomic Units (OTUs) obtained for each sample are shown. The mean quality score (Q score), the percentage of sequences with a Q score higher than 30 (Q30) and the Shannon biodiversity index for each sequencing run have also been included. Sample names: PC, perimeter channel; PZ, piezometer; TR, Tinto River; SW, Sea water.

712 **REFERENCES**

- Agarwal, M., Rathore, R. S., Jagoe, C., & Chauhan, A. (2019). Multiple Lines of
 Evidences Reveal Mechanisms Underpinning Mercury Resistance and
 Volatilization by Stenotrophomonas sp. MA5 Isolated from the Savannah River
 Site (SRS), USA. Cells, 8(4), 309. https://doi.org/10.3390/cells8040309
- Aguilera, A. (2013). Eukaryotic Organisms in Extreme Acidic Environments, the Río Tinto
 Case. Life, 3(3), 363–374. https://doi.org/10.3390/life3030363
- Auld, R. R., Myre, M., Mykytczuk, N. C. S., Leduc, L. G., & Merritt, T. J. S. (2013).
 Characterization of the microbial acid mine drainage microbial community using
 culturing and direct sequencing techniques. Journal of Microbiological Methods,
 93(2), 108–115. https://doi.org/10.1016/j.mimet.2013.01.023
- Aytar, P., Kay, C. M., Mutlu, M. B., Çabuk, A., & Johnson, D. B. (2014). Diversity of
 acidophilic prokaryotes at two acid mine drainage sites in Turkey. Environmental
 Science and Pollution Research, 22(8), 5995–6003.
 <u>https://doi.org/10.1007/s11356-014-3789-4</u>
- Bilgin, A. A., Silverstein, J., & Jenkins, J. D. (2004). Iron respiration by Acidiphilium
 cryptum at pH 5. FEMS Microbiology Ecology, 49(1), 137–143.
 https://doi.org/10.1016/j.femsec.2003.08.018
- Bohorquez, L. C., Delgado-Serrano, L., López, G., Osorio-Forero, C., Klepac-Ceraj, V.,
 Kolter, R., Junca, H., Baena, S., & Zambrano, M. M. (2011). In-depth
 Characterization via Complementing Culture-Independent Approaches of the
 Microbial Community in an Acidic Hot Spring of the Colombian Andes. Microbial
 Ecology, 63(1), 103–115. https://doi.org/10.1007/s00248-011-9943-3.
- Bolyen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C. C., Al-Ghalith, G. A.,
 Alexander, H., Alm, E. J., Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J. E.,
 Bittinger, K., Brejnrod, A., Brislawn, C. J., Brown, C. T., Callahan, B. J., CaraballoRodríguez, A. M., Chase, J., ... Caporaso, J. G. (2019). Reproducible, interactive,
 scalable and extensible microbiome data science using QIIME 2. Nature
 Biotechnology, 37(8), 852–857. https://doi.org/10.1038/s41587-019-0209-9
- Bolívar, J. P., Martín, J. E., García-Tenorio, R., Pérez-Moreno, J. P., & Mas, J. L. (2009).
 Behaviour and fluxes of natural radionuclides in the production process of a
 phosphoric acid plant. Applied Radiation and Isotopes, 67(2), 345–356.
 https://doi.org/10.1016/j.apradiso.2008.10.012
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes,
 S. P. (2016). DADA2: High-resolution sample inference from Illumina amplicon
 data. Nature Methods, 13(7), 581–583. https://doi.org/10.1038/nmeth.3869
- De Vos, W., & Tarvainen, T. (2006). The Geochemical Atlas of Europe Part 2.
 Interpretation of Geochemical Maps, Additional Tables, Figures, Maps, and
 Related Publications. 690 pp.
- Douglas, G. M., Maffei, V. J., Zaneveld, J. R., Yurgel, S. N., Brown, J. R., Taylor, C. M.,
 Huttenhower, C., & Langille, M. G. I. (2020). PICRUSt2 for prediction of
 metagenome functions. Nature Biotechnology, 38(6), 685–688.
 https://doi.org/10.1038/s41587-020-0548-6

- Jersanek, C. D., Weithoff, G., & Weisse, T. (2011). *Cephalodella acidophila* n. sp.
 (Monogononta: Notommatidae), a new rotifer species from highly acidic mining
 lakes. Zootaxa, 2939(1), 50. https://doi.org/10.11646/zootaxa.2939.1.2
- Gadanho, M., Libkind, D., & Sampaio, J. P. (2006). Yeast Diversity in the Extreme Acidic
 Environments of the Iberian Pyrite Belt. Microbial Ecology, 52(3), 552–563.
 https://doi.org/10.1007/s00248-006-9027-y
- Gavrilov, S. N., Korzhenkov, A. A., Kublanov, I. V., Bargiela, R., Zamana, L. V., Popova,
 A. A., Toshchakov, S. V., Golyshin, P. N., & Golyshina, O. V. (2019). Microbial
 Communities of Polymetallic Deposits' Acidic Ecosystems of Continental Climatic
 Zone with High Temperature Contrasts. Frontiers in Microbiology, 10.
 https://doi.org/10.3389/fmicb.2019.01573
- Ghosh, W., & Dam, B. (2009). Biochemistry and molecular biology of lithotrophic sulfur
 oxidation by taxonomically and ecologically diverse bacteria and archaea. FEMS
 Microbiology Reviews, 33(6), 999–1043. https://doi.org/10.1111/j.1574 6976.2009.00187.x
- González-Toril, E., Llobet-Brossa, E., Casamayor, E. O., Amann, R., & Amils, R. (2003).
 Microbial Ecology of an Extreme Acidic Environment, the Tinto River. Applied and
 Environmental Microbiology, 69(8), 4853–4865.
 https://doi.org/10.1128/aem.69.8.4853-4865.2003
- Gómez-Villegas, P., Vigara, J., & León, R. (2018). Characterization of the microbial
 population inhabiting a solar saltern pond of the odiel Marshlands (SW Spain).
 Marine Drugs, *16*(9), 332. doi: 10.3390/md16090332
- Guerrero, J. L., Gutiérrez-Álvarez, I., Mosqueda, F., Gázquez, M. J., García-Tenorio, R.,
 Olías, M., & Bolívar, J. P. (2020). Evaluation of the radioactive pollution in the
 salt-marshes under a phosphogypsum stack system. Environmental Pollution,
 258, 113729. https://doi.org/10.1016/j.envpol.2019.113729
- Guerrero, J. L., Pérez-Moreno, S. M., Gutiérrez-Álvarez, I., Gázquez, M. J., & Bolívar, J.
 P. (2021a). Behaviour of heavy metals and natural radionuclides in the mixing of
 phosphogypsum leachates with seawater. Environmental Pollution, 268, 115843.
 https://doi.org/10.1016/j.envpol.2020.115843
- Guerrero, J. L., Gutiérrez-Álvarez, I., Hierro, A., Pérez-Moreno, S. M., Olías, M., &
 Bolívar, J. P. (2021b). Seasonal evolution of natural radionuclides in two rivers
 affected by acid mine drainage and phosphogypsum pollution. CATENA, 197,
 104978. https://doi.org/10.1016/j.catena.2020.104978
- Hou, D., Zhang, P., Wei, D., Zhang, J., Yan, B., Cao, L., Zhou, Y., & Luo, L. (2020).
 Simultaneous removal of iron and manganese from acid mine drainage by
 acclimated bacteria. Journal of Hazardous Materials, 396, 122631.
 https://doi.org/10.1016/j.jhazmat.2020.122631
- Kisková, J., Perháčová, Z., Vlčko, L., Sedláková, J., Kvasnová, S., & Pristaš, P. (2018).
 The Bacterial Population of Neutral Mine Drainage Water of Elizabeth's Shaft
 (Slovinky, Slovakia). Current Microbiology, 75(8), 988–996.
 https://doi.org/10.1007/s00284-018-1472-6
- León-Vaz, A., Romero, L. C., Gotor, C., León, R., & Vigara, J. (2021). Effect of cadmium
 in the microalga Chlorella sorokiniana: A proteomic study. Ecotoxicology and

 799
 Environmental
 Safety,
 207,

 800
 https://doi.org/10.1016/j.ecoenv.2020.111301

- Li, L., Liu, Z., Zhang, M., Meng, D., Liu, X., Wang, P., Li, X., Jiang, Z., Zhong, S., Jiang,
 C., & Yin, H. (2020). Insights into the Metabolism and Evolution of the Genus
 Acidiphilium, a Typical Acidophile in Acid Mine Drainage. MSystems, 5(6).
 https://doi.org/10.1128/msystems.00867-20
- López-Archilla, A. I., Marin, I., & Amils, R. (2001). Microbial Community Composition and
 Ecology of an Acidic Aquatic Environment: The Tinto River, Spain. Microbial
 Ecology, 41(1), 20–35. https://doi.org/10.1007/s002480000044
- López-Archilla, A. I., González, A. E., Terrón, M. C., & Amils, R. (2004). Ecological study
 of the fungal populations of the acidic Tinto River in southwestern Spain.
 Canadian Journal of Microbiology, 50(11), 923–934. https://doi.org/10.1139/w04089
- Lukhele, T., Selvarajan, R., Nyoni, H., Mamba, B. B., & Msagati, T. A. M. (2020). Acid
 Mine Drainage as Habitats for Distinct Microbiomes: Current Knowledge in the
 Era of Molecular and Omic Technologies. Current Microbiology, 77(4), 657–674.
 https://doi.org/10.1007/s00284-019-01771-z
- Mefteh, B., Bouket, Daoud, Luptakova, Alenezi, Gharsallah, & Belbahri. (2019).
 Metagenomic Insights and Genomic Analysis of Phosphogypsum and Its
 Associated Plant Endophytic Microbiomes Reveals Valuable Actors for Waste
 Bioremediation. Microorganisms, 7(10), 382.
 https://doi.org/10.3390/microorganisms7100382
- Méndez-García, C., Peláez, A. I., Mesa, V., Sánchez, J., Golyshina, O. V., & Ferrer, M.
 (2015). Microbial diversity and metabolic networks in acid mine drainage habitats.
 Frontiers in Microbiology, 6. https://doi.org/10.3389/fmicb.2015.00475
- Merino, N., Aronson, H. S., Bojanova, D. P., Feyhl-Buska, J., Wong, M. L., Zhang, S., & 824 825 Giovannelli, D. (2019). Living at the Extremes: Extremophiles and the Limits of Life in Planetary 826 а Context. Frontiers in Microbiology, 10. https://doi.org/10.3389/fmicb.2019.00780 827
- OSPAR, (2002). Discharges of Radioactive Substances into the Maritime Area by
 Nonnuclear Industry. Radioactive Substances Series. Publication No. 161.
 OSPAR Commission, London, p. 60.
- Papaslioti, E.-M., Pérez-López, R., Parviainen, A., Sarmiento, A. M., Nieto, J. M.,
 Marchesi, C., Delgado-Huertas, A., & Garrido, C. J. (2018). Effects of seawater
 mixing on the mobility of trace elements in acid phosphogypsum leachates.
 Marine Pollution Bulletin, 127, 695–703.
 https://doi.org/10.1016/j.marpolbul.2018.01.001
- Pérez-López, R., Álvarez-Valero, A. M., & Nieto, J. M. (2007). Changes in mobility of
 toxic elements during the production of phosphoric acid in the fertilizer industry
 of Huelva (SW Spain) and environmental impact of phosphogypsum wastes.
 Journal of Hazardous Materials, 148(3), 745–750.
 https://doi.org/10.1016/j.jhazmat.2007.06.068
- Pérez-López, R., Macías, F., Cánovas, C. R., Sarmiento, A. M., & Pérez-Moreno, S. M.
 (2016). Pollutant flows from a phosphogypsum disposal area to an estuarine

- 843 environment: An insight from geochemical signatures. Science of The Total 844 Environment, 553, 42–51. https://doi.org/10.1016/j.scitotenv.2016.02.070
- Pérez-López, R., Carrero, S., Cruz-Hernández, P., Asta, M. P., Macías, F., Cánovas, C.
 R., Guglieri, C., & Nieto, J. M. (2018). Sulfate reduction processes in salt marshes
 affected by phosphogypsum: Geochemical influences on contaminant mobility.
 Journal of Hazardous Materials, 350, 154–161.
 https://doi.org/10.1016/j.jhazmat.2018.02.001
- Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J., &
 Glöckner, F. O. (2012). The SILVA ribosomal RNA gene database project:
 improved data processing and web-based tools. Nucleic Acids Research, 41(D1),
 D590–D596. https://doi.org/10.1093/nar/gks1219
- Rognes, T., Flouri, T., Nichols, B., Quince, C., & Mahé, F. (2016). VSEARCH: a versatile
 open source tool for metagenomics. PeerJ, 4, e2584.
 https://doi.org/10.7717/peerj.2584
- Russo, G., Libkind, D., Sampaio, J. P., & Van Broock, M. R. (2008). Yeast diversity in
 the acidic Rio Agrioâ[□] "Lake Caviahue volcanic environment (Patagonia,
 Argentina). FEMS Microbiology Ecology, 65(3), 415–424.
 https://doi.org/10.1111/j.1574-6941.2008.00514.x
- 861 Santofimia, E., González-Toril, E., López-Pamo, E., Gomariz, M., Amils, R., & Aguilera, 862 Á. (2013). Microbial Diversity and Its Relationship to Physicochemical 863 Characteristics of the Water in Two Extreme Acidic Pit Lakes from the Iberian e66746. 864 Pvrite Belt (SW Spain). PLoS ONE, 8(6), https://doi.org/10.1371/journal.pone.0066746 865
- Trifi, H., Najjari, A., Achouak, W., Barakat, M., Ghedira, K., Mrad, F., Saidi, M., & Sghaier,
 H. (2020). Metataxonomics of Tunisian phosphogypsum based on five
 bioinformatics pipelines: Insights for bioremediation. Genomics, 112(1), 981–989.
 https://doi.org/10.1016/j.ygeno.2019.06.014
- Zettler, L. A. A., Gómez, F., Zettler, E., Keenan, B. G., Amils, R., & Sogin, M. L. (2002).
 Eukaryotic diversity in Spain's River of Fire. Nature, 417(6885), 137–137.
 https://doi.org/10.1038/417137a
- Zouch, H., Karray, F., Armougom, F., Chifflet, S., Hirschler-Réa, A., Kharrat, H., Kamoun,
 L., Ben Hania, W., Ollivier, B., Sayadi, S., & Quéméneur, M. (2017). Microbial
 Diversity in Sulfate-Reducing Marine Sediment Enrichment Cultures Associated
 with Anaerobic Biotransformation of Coastal Stockpiled Phosphogypsum (Sfax,
 Tunisia). Frontiers in Microbiology, 8. https://doi.org/10.3389/fmicb.2017.01583









MOST ABUNDANT EUKARYOTA PHYLA (%)









PZ

в PC RT 3 2 3 390 5 14 10 SW PZ RT 5 1 393 5 11 10 SW PC 399 11 11

PZ

С

PC-RT-SW

UNIQUE TO THE PC: PWY-4722; PWY-5656 UNIQUE TO THE PC+RT: P341-PWY; PWY-6142; PWY-3081 UNIQUE TO THE PC+SW: PWY-6948; PWY-6404; PWY-7373; PWY-6581; PWY-1422

PZ-RT-SW

UNIQUE TO THE P2: PWY-5656 UNIQUE TO THE P2+RT: P341-PWY UNIQUE TO THE P2+SW: PWY-6948; PWY-6404; PWY-7373; PWY-7398; PWY-6143

PZ-PC

UNIQUE TO THE PZ: PWY-4722; PWY-6142; PWY-5177; LPSYN-PWY; THREOCAT-PWY; PWY-6339; PWY-658115; PWY-1422; P162-PWY; PWY-1361; PWY-3081

UNIQUE TO THE PC: PWY-6167; PWY-6349; PWY-7024; PWY-7046; PWY-5743; P241-PWY; PWY-5744; PWY-6654; PWY-6350; PWY-739815/09/2021 PWY-6143